



COMMENTARY

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Suitable test substances for proof of concept regarding infochemical effects in surface waters

Monika Nendza¹, Ursula Klaschka^{2*} and Rüdiger Berghahn³**Abstract**

Background: Infochemical effects have been defined as the manipulation of the odour perception of organisms by anthropogenic substances which may result in ecologically relevant behavioural disorder. However, the environmental relevance of infochemical effects has not yet been confirmed by experimental observations. This project aims to test for infochemical effects on chemical communication in water bodies with systematic experimental investigations. The first crucial step is to select suitable test substances. Repellents (PT 19 biocides) and odourants may be assumed to affect the response of aquatic populations and communities. These mostly polar and stable compounds may disturb chemical communication between organisms and may cause organismic effects like drift (downstream dislocation of e.g. crustacean and insect larvae in streams). Repellents enter surface waters mainly indirectly via wastewater discharges from sewage treatment plants or directly by being washed off from the skin and clothes of bathers.

Results: In this literature study, suitable chemicals were selected for confirmatory assessments of suspected infochemical effects by laboratory tests in a subsequent second part of the project. The use pattern and physico-chemical properties of the substances selected, in combination with their limited biological degradability, indicate potential aquatic relevance with possible chronic impact caused by disturbed communication. After due consideration of advantages and limitations, three PT 19 repellents appear suitable test compounds for proof of concept in the subsequent experimental part of the project:

- DEET (CAS 134-62-3)
- Icaridine (CAS 119515-38-7)
- EBAAP (CAS 52304-36-6)

Another promising candidate for infochemical effects is isophorone (CAS 78-59-1), a natural attractant and an anthropogenic high production volume solvent.

Conclusions: Four chemicals were selected with the expectation that they may be suitable test substances for experimental proof of concept of infochemical effects in the subsequent part of the project. The experimental results may then help to answer the question of whether PT 19 biocides and other odourants entering aquatic ecosystems give rise for concern about potential infochemical effects.

Keywords: Biocides; Chemical communication; Drift; Infochemicals; Non-target organisms; Odourants; Repellents; Surface water

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Background

The infochemical effect: misled by pollutants?

Most organisms live in an odourous environment and perceive their biotic and abiotic environment via specific and dynamic blends of odourants called infochemicals [1-6]. They are defined as 'a chemical that carries information that mediates an interaction among two individuals and results in an adaptive response in the receiver. Either the sender or the receiver, or both, benefit from the infochemical' [1]. Infochemicals play an important role in life history (e.g. such as mate choice, voltinism, generation time, clutch size), habitat finding, food recognition and survival (e.g. response to predation threats). Infochemicals are major means of communication in aquatic ecosystems because other senses, e.g. vision and mechanical senses, may become less efficient in nature under turbid and turbulent conditions [1-7]. Clouds of odours from various sources can overlap and lead to different perceptions. A comprehensive compilation of current knowledge about infochemicals in aquatic ecosystems was published by Brönmark and Hansson in 2012 [7]. However, as has been demonstrated for various invertebrates and fish, this sensitive system for communication and detection may be disturbed by discharges of pollutants such as fragrances, metals or pesticides [8]. This disturbance is referred to as the infochemical effect [8] or infodisruption [9,10]. The consideration of infochemical effects of anthropogenic substances means transferring knowledge gained in chemical ecology to the field of ecotoxicology.

The current testing strategies for ecotoxicological endpoints do not consider effects on chemical communication at the population and biocoenosis level. Effective concentrations of anthropogenic infochemicals are in the range of micrograms per litre [11]. Infochemicals are mostly released in dynamic gradients at sites where the local concentrations of microenvironments are relevant for the receiver of a chemical cue [11,12].

Ecological relevance of infochemical effects

Since chemical communication is very important for aquatic organisms, it is likely that the disturbance of this sensitive system may have effects on populations and ecosystems. However, experimental evidence of infochemical effects of anthropogenic substances in surface waters is weak as yet. Testing for infochemical effects in the laboratory is a challenge for two reasons: (1) suitable test chemicals which are true positives need to be anticipated and (2) test systems must be adapted to discriminate infochemical effects from other subacute toxicities like avoidance reactions. A systematic analysis will help to answer the question whether and to what extent anthropogenic chemicals may interfere with chemical communication in natural ecosystems [13].

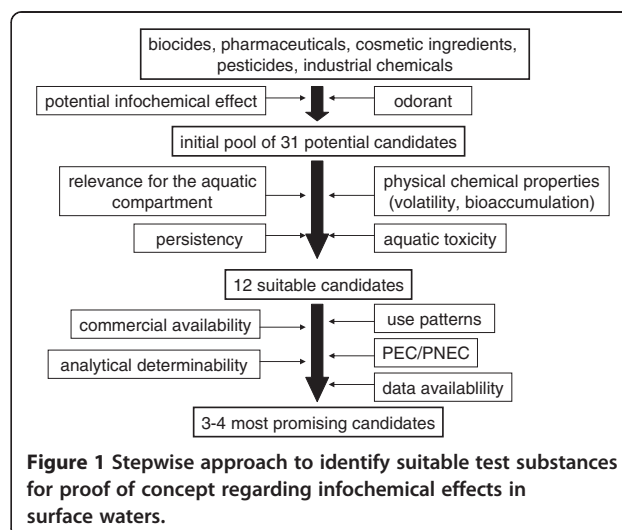
Testing strategy

The number of potential anthropogenic infochemicals is large, as is the number of test designs [7,13,14]. Since experimental testing is time-consuming and costly, we took great care beforehand to select promising test substances with potential infochemical effect. We started with repellents (biocides product type (PT) 19) for which smell is the primary mode of action. Furthermore, we focussed on odourants used in e.g. cosmetics and industrial chemicals. In the second part of this project, the selected candidate substances will be subjected to already established behavioural assays addressing possible infochemical effects, namely, vertical migration of daphnids, aggregation of aquatic pulmonates and organismic drift in artificial streams [14-18].

Selection of test substances

The search for potent infochemicals started with the large number of chemicals that can possibly interfere with odour perception in water [7,8,13]. We used a stepwise approach to identify suitable test compounds among olfactory receptor-binding substances (Figure 1). Realistic infochemical potential can be assumed for three groups of substances:

1. Repellents and attractants belonging to product group 19 (PT 19) according to the Biocides Regulation (1451/2007) [19]. These substances are used to keep away or attract target organisms, e.g. invertebrates such as midges or fleas or vertebrates such as birds or boars. The common mode of action is the smelling of the active ingredients. Repellents and attractants are usually non-toxic to the target organisms. They can be natural compounds, e.g. essential oils, or synthetic compounds, such as N,N-diethyl-3-methylbenzamide (DEET), icaridine



or ethyl-N-acetyl-N-butyl- β -alaninate (EBAAP). It is likely that repellents and attractants can influence also the behaviour and communication of non-target organisms. In the risk assessment of repellents, specific effects on non-target organisms in surface waters are not yet considered. We hypothesize that (1) repellents are potential candidates of anthropogenic infochemicals since their mode of action aims at misleading the chemical communication of the target organisms. Thus, (2) repellents may be involved in unwanted effects such as organismic drift of non-target species. Major effects on downstream drift [18] are especially relevant for animals that pupate or hatch only once per year or even after multiannual larval stages (e.g. dragonflies or some caddy fly species). Catastrophic downstream drift may strongly affect the size and the recovery potential of populations [20,21]. Based on current knowledge, it is impossible to say whether and to what extent organismic drift may be induced by infochemical effects of PT 19 biocides at relevant field concentrations. The following questions summarize the issue: may repellents disturb communication between aquatic non-target organisms? May this affect the structure (e.g. biodiversity) and function (e.g. predator-prey interactions) of ecosystems?

2. Natural infochemicals are substances that, for example, signal sites for egg deposition or food availability, alarm substances and pheromones. The chemical cues are usually blends of compounds released by organisms. Only few natural infochemicals have been chemically identified, and clear concentration-effect relationships for individual substances are rarely described. Natural infochemicals can be positive controls of test setup, procedures and results. Furthermore, structural similarities between natural infochemicals and anthropogenic substances may support the identification of suitable test substances.
3. Potential anthropogenic infochemicals may also be found among cosmetic ingredients, pesticides, pharmaceuticals and industrial chemicals that affect the olfactory reaction cascade. Fragrances used in scented products are further candidates for the infochemical effect; some of them are even chemically identical with natural infochemicals [8,22]. Anthropogenic infochemicals may disturb chemical communication by imitating or modulating natural cues.

We conducted a literature study in order to find out, based on available information on quantities used, fate and effects, which substances out of these three groups

might be relevant candidates for the testing of infochemical effects in the aquatic environment.

Methods

The literature and database searches for the identification of candidate substances drew on publicly available literature, software and databases. The list of repellents and attractants belonging to product group 19 (PT 19) was taken from the Biocides Regulation (1451/2007) [19] and its implementing rules in Commission Regulations 1048 (2005) and 1849 (2006). Lists of natural and anthropogenic substances were taken from the literature [5,10,22,23]. Assessments of exposure to the substances and their relevance for the water compartment were performed using monitoring data (concentrations and loads in surface water), considering different entry paths (e.g. direct and indirect discharge via water and air, runoff, drainage, wastewater treatment plants, rainwater retention, bathers) and the distribution of the substances in the environment (percentage in water, sediment, soil, air [24]).

Inherent properties of the substances relevant to exposure (water solubility, log *K*_{ow}, Henry constant, persistence) as well as indications of aquatic toxicities (EC/LC₅₀, *predicted no-effect concentration* (PNEC) values, classification and labelling, e.g. as R50/53) were compiled from multiple sources [24-33]. Monitoring data were taken from [25,27,28,30,34-44]. In cases where experimental information was insufficient, estimations were made using EpiSuite 4.1 [24].

The aquatic exposure assessments made here should exclusively be used for the prioritization of candidate substances and are not intended for regular risk assessments.

The complete data sets are open to the public in the UBA report FKZ 3712 67 417.1 'Relevance of effects of repellents (product type 19) and other infochemicals for non-target organisms in surface waters, part I: literature study' (in German: Wirkungsrelevanz von Repellentien (Produktart 19) und anderen Infochemikalien für Nichtzielorganismen in Oberflächengewässern, Teil I: Literaturstudie).

Discussion

In the first selection step (Figure 1), we started with several pragmatic criteria to focus on possible test substances:

- Repellents were preferred over substances with other modes of action.
- Substances without direct toxicities at low doses were preferred to rule out the possibility that other reactions overrule the infochemical effects. Hence, PT 18 substances (insecticides, acaricides and products against other arthropods) and chemicals

subject to classification and labelling (R50-R53, H400-H412) or showing other indications of environmental toxicity were excluded.

- Substances used as repellents against insects were preferred to substances used as repellents against vertebrates, as the non-target organisms in the aquatic environment are mainly invertebrates.
- Single substances were preferred to mixtures of natural substances for two reasons. In the case of natural substances, monitoring data may be obscured by background concentrations, and experimental tests with single substances are more likely to produce clear concentration-response curves.
- Substances discharged into surface waters via wastewater or bathers by being washed off from human skin or clothes were preferred to substances emitted to air or soil, since the focus of this study was on aquatic non-target organisms.

This first selection step was very efficient and yielded 31 compounds, which are listed with their CAS numbers below:

- PT 19 biocides: icaridine (119515-38-7), EBAAP (52304-36-6), DEET (134-62-3), citriodiol (42822-86-6), lauric acid (143-07-7), linalool (78-70-6), lavender (91722-69-9), pelargonic acid (112-05-0), methyl-n-nonylketone (112-12-9), methyl anthranilate (134-20-3), garlic extract (8008-99-9), geraniol (106-24-1), naphthalene (91-20-3), margosa extract (84696-25-3), pyrethrins and pyrethroids (8003-34-7), n-decanoic acid (334-48-5), carbon dioxide (124-38-9), 1-octen-3-ol (3391-86-4), cis-9-tricosene (27519-02-4), trimethyloctanamide (105726-67-8), 9,12-tetradecadienyl acetate (30507-70-1), piperonyl butoxide (51-03-6).
- Natural infochemicals: microcystin LR (101043-37-2), hypoxanthine-3-N-oxide (19765-65-2), isophorone (78-59-1).
- Anthropogenic infochemicals: metalochlor (51218-45-2), copper compounds (various speciations), carbaryl (63-25-2), D-limonene (5989-27-5), benzaldehyde (100-52-7), tridecanone (593-08-8).

The criteria applied in the second selection step (Figure 1) focussed on physico-chemical properties of the selected substances and preliminary screening for exposure indicating relevant concentrations in aquatic compartments. The intention was to identify a small subset of substances with possible relevance for surface waters:

- Substances used in consumer products or detected in surface waters.

- Substances not likely to be phased out in the near future.
- Substances with substantial water solubility (>1 mg/L).
- Substances for which exposure modelling indicates that at least 20% of the amount discharged into the environment ends up in the water compartment.
- Substances of low volatility from the water compartment (small Henry constant).
- Substances of medium lipophilicity ($\log K_{ow} < 4$), indicating minor to medium bioaccumulation potential and favourable properties for the conduct of laboratory tests (i.e. low adsorption to test vessels and appliances).
- Substances with sufficient stability to ensure uniform concentrations during laboratory tests.

Twelve substances (Figure 2) mostly complied with the criteria in this second prioritization step.

In the third selection step, we identified the most promising test substances (Table 1) after a ranking of the 12 candidate substances with regard to:

- Infochemical potential.
- Relevance to aquatic compartments.
- Minor direct toxicities at low doses.
- Physico-chemical suitability for testing.

Available data for the 12 compounds were used to evaluate their persistence and relevance for aquatic ecosystems, their bioavailability and aquatic toxicities, as well as their technical amenability to laboratory testing including analytical methods, commercial availability of testing material and analytical standards. These considerations are discussed here in more detail (see Table 1 for the relevant data):

Most insect repellents enter the environment indirectly via wastewater and sewage treatment plants and directly via bathing in surface waters (by being washed off from skin and clothes). In the case of DEET, 95% of the substance is discharged via municipal wastewater treatment plants [45]. No information was found about the amounts reaching the environment through direct inputs from leisure activities. DEET was detected in all analyzed influents and effluents of wastewater treatment plants in Germany as well as in surface waters used for human activities [25]. Concentrations of up to 0.6 µg/L were detected during holiday times in winter and spring, but they were much lower than those in summer. Correspondingly, it was assumed for other insect repellents that the discharge is highest in summer, between June and September. DEET removal rates in sewage treatment plants varied between 0% and 90% depending on the concentrations in the influents, probably due to the

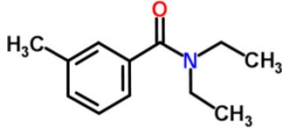
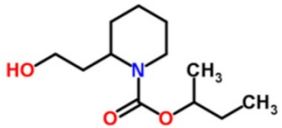
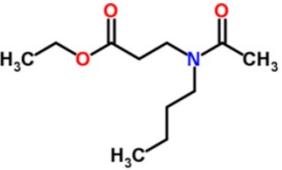
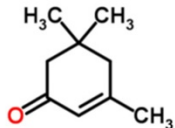
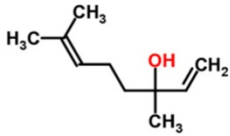
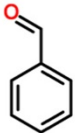
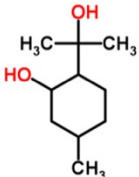
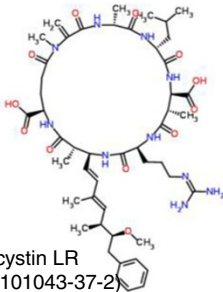
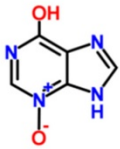
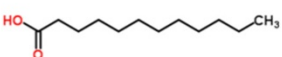
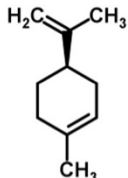
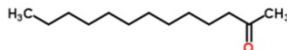
			
DEET (CAS 134-62-3)	Icaridine (CAS 119515-38-7)	EBAAP (CAS 52304-36-6)	Isophorone (CAS 78-59-1)
			
Linalool (CAS 78-70-6)	Benzaldehyde (CAS 100-52-7)	Citriodiol (CAS 42822-86-6)	Microcystin LR (CAS 101043-37-2)
			
Hypoxanthine-3-N-oxide (CAS 19765-65-2)	Lauric acid (CAS 143-07-7)	D-Limonene (CAS 5989-27-5)	Tridecanone (CAS 593-08-8)

Figure 2 Chemical structures and CAS numbers of 12 candidate substances. For the experimental assessment of infochemical effects as a result of selection step 2.

adaptation of the sludge organisms. Icaridine is characterized as a readily degradable substance and was detected in influents but not in effluents of sewage treatment plants [25].

According to the Canadian Categorization Results [26], water is an important environmental compartment for all 12 substances selected in step 2 (Figure 2). Exceptions are D-limonene and tridecanone. Calculations using the level III fugacity model of EpiSuite 4.1 [24] indicated that the candidate substances partition into the water phase at rates of about 20% or more (Table 1). The only remarkable exception is the water-soluble natural algal toxin microcystin LR.

Monitoring data are available only for some of the candidate substances. For most of these, measured concentrations in water bodies indicate a low risk compared with published PNEC values. For DEET, measured concentrations in North American surface waters are in the same range as the standard PNEC values of 0.043 [27] or 0.076 mg/L [28] and indicate a potential risk for US

waters. Concentrations detected in Europe are about one order of magnitude lower (Table 1). Since 1998, the readily degradable icaridine (Bayrepel®) has been used as substitute for DEET and, as expected, its monitoring values with negligible concentrations in STP effluents reflect the ready degradability in sewage treatment plants [25]. No monitoring data for German or European surface waters could be found for EBAAP (IR3535®), an agent in modern consumer products. It is worthwhile to mention that some monitoring data on PT19 biocides and other contaminants were said to be biased due to contamination of samples by insect repellents applied by field workers during sampling campaigns [46]. The detection of DEET in field blanks as well as the contamination of reagents with isophorone prompted the authors [46] to exclude the respective values.

The assessment of the stability of the candidate substances was not straightforward. The available information on the biological degradability of the candidate substances (Table 1) is heterogeneous and sometimes

Table 1 Properties of promising candidate substances for the experimental assessment of infochemical effects

	DEET	Reference	Icaridine	Reference	EBAAP	Reference	Isophorone	Reference
CAS	134-62-3		119515-38-7		52304-36-6		78-59-1	
Infochemical potential	Insect repellent, PT19 biocide	[19]	Insect repellent, PT19 biocide	[19]	Insect repellent, PT19 biocide	[19]	Natural attractant and anthropogenic HPV solvent (host recognition)	[5]
Water solubility	11.2 g/L	[29]	8.2 g/L	[29]	70 g/L	[29]	14.5 g/L	[30]
			10.6 g/L	[29]			12 g/L	[30]
Log K _{ow}	2.4	[29]	2.23	[29]	1.7	[29]	1.67	[30]
	2.18	[26]	2.57	[24]	1.51 (calculated)	[24]	1.66	[30]
							1.73	[30]
Henry constant	3.93×10^{-3}	[29]	9.1×10^{-4}	[29]	4.61×10^{-4} (calculated)	[29]	6.73×10^{-1}	[24]
	2.10×10^{-3} (calculated)	[24]	3.01×10^{-6} (calculated)	[24]	5.36×10^{-5} (calculated)	[24]		
Degradability	Not persistent	[29]	Potentially persistent	[29]	Not persistent	[29]	Persistent	[26]
	Persistent	[26]						
	Readily biodegradable	[29]	Not readily biodegradable	[29]	Not readily biodegradable	[29]	95% (readily) biodegradable	[30]
	0% biodegradable	[26]	Not inherently biodegradable	[29]	Readily biodegradable (calculated)	[24]	Inherently biodegradable	[30]
	Not biodegradable	[28]	Readily biodegradable	[28]			1.5% biodegradable	[26]
	Not readily biodegradable (calculated)	[24]	Not readily biodegradable (calculated)	[24]			Not readily biodegradable (calculated)	[24]
	Minor hydrolysis	[29]	Hydrolytically stable	[29]	Minor hydrolysis	[29]	Hydrolysis not relevant	[30]
Calculated compartmental distribution with the level III fugacity model of EpiSuite 4.1		[24]						
Water	18.6%		20.1%		23.5%		27.4%	
Sediment	0.137%		0.102%		0.078%		0.137%	
Soil	81.1%		79.8%		76.4%		72.4%	
Air	0.126%		0.00056%		0.0101%		0.109%	
Monitoring	0.6 to 2.33 µg/L (STP effluent)	[25]	0.6 to 1 µg/L (STP influent) < 0.1 µg/L (STP effluent)	[25]			<0.5 µg/L (detection limit)	[39]
	0.02 to 1.13 µg/L (stream)	[42]	<0.1 to 2.2 µg/L	[44]			<0.1 µg/L (background), 10 µg/L (exposure)	[30]
	8 to 3000 ng/L	[37]	0 to 0.01 µg/L	[41]			0.65 to 9.1 ng/L (lake)	[34]
	≤3 µg/L	[28]						
	<0.08 to 6.9 µg/L	[44]						

Table 1 Properties of promising candidate substances for the experimental assessment of infochemical effects (Continued)

	<0.4 to 454 ng/L (ground water)	[40]					
	0.06 µg/L (surface water)	[43]					
	0.03 µg/L (drinking water)						
	0.05 to 20 µg/L (STP influent)	[35]					
	0.02 to 15 µg/L (STP effluent)						
	0.007 to 33.4 µg/L (surface water USA)						
	0.002 to 1.3 µg/L (surface water)						
	13 to 660 ng/L (surface water, median of 188 samples; 55 ng/L)	[36]					
	Approximately 0.03 µg/L (range, 0.005 to 0.2 µg/L)	[38]					
	0 to 0.1 µg/L	[41]					
	29 to 52 ng/L (lake)	[34]					
	64 to 245 ng/L	[27]					
PNEC	0.043 mg/L	[27]	0.31 mg/L	[29]	Not toxic to the aquatic environment	[29]	0.089 mg/L [30]
	0.076 mg/L	[28]					
Aquatic toxicities	96 h LC50 97 mg/L (<i>Danio rerio</i>)	[29]	96 h LC50 169.4 mg/L (<i>Oncorhynchus mykiss</i>)	[29]	96 h LC50 >100 mg/L (<i>Danio rerio</i>)	[29]	96 h LC50 140 mg/L (<i>Cyprinodon variegatus</i>) [30]
	96 h LC50 110 mg/L (<i>Pimephales promelas</i>)	[32]	32 d NOEC 3.14 mg/L (<i>Danio rerio</i>)	[29]	48 h LC50 >100 mg/L (<i>Daphnia magna</i>)	[29]	96 h EC50 217 mg/L, LC50 228 mg/L (<i>Pimephales promelas</i>) [30]
	96 h LC50 71 mg/L (<i>Oncorhynchus mykiss</i>)	[33]	48 h LC50 >103 mg/L (<i>Daphnia magna</i>)	[29]	72 h LC50 >100 mg/L (<i>Desmodesmus subspicatus</i>)	[29]	32 d NOEC 4.2 mg/L (<i>Pimephales promelas</i>) [30]
	51 h LC50 75 mg/L (<i>Daphnia magna</i>)	[29]	21 d NOEC 49.25 mg/L (<i>Daphnia magna</i>)	[29]			35 d NOEC 11 mg/L, LOEC 19 mg/L (<i>Pimephales promelas</i>) [30]
	96 h LC50 100 mg/L (<i>Gammarus fasciatus</i>)	[31]	72 h ErC50 87.3 mg/L, NOEC 54.8 mg/L (<i>Scenedesmus subspicatus</i>)	[29]			48 h NOEC 15 mg/L, LC50 120 mg/L (<i>Daphnia magna</i>) [30]
							72 h EC10 64 mg/L, EC50 475 mg/L (<i>Scenedesmus subspicatus</i>) [30]

Properties of promising candidate substances for the experimental assessment of infochemical effects due to infochemical potential, relevance to aquatic compartments, minor direct toxicities at low doses and physico-chemical liability for testing. The data are experimentally determined unless explicitly marked (calculated) [24].

even contradictory. Evidence indicates that DEET [28,29] and isophorone [26,30] are likely more persistent than icaridine [28,29] and EBAAP [29]. Pseudo-persistence may be relevant for some of the candidates which are discharged continuously into surface waters. All candidate substances, but benzaldehyde, do not hydrolyze or hydrolyze slowly and are therefore sufficiently stable in laboratory tests.

Experimental data on aquatic toxicity was found for all candidate substances, with the exception of citriodiol. Missing data for hypoxanthine-3-N-oxide, microcystin LR and tridecanone were supplemented by calculated values [24]. The toxicity values reported are usually results from conventional tests. The only quantitative result for an infochemical effect was found for hypoxanthine-3-N-oxide with short-term exposure (7 min) leading to a concentration-dependent alarm reaction in zebra fish [47]. As expected, erratic movements and jumps occurred at low concentrations (LOEC 0.23 mg/L [47]) while conventional endpoints are likely affected at much higher levels (LC50 >100 mg/L [24]). However, measured LC50 data for hypoxanthine-3-N-oxide could not be retrieved, and the comparison of estimated effect levels is subject to major uncertainties. Data on long-term effects could only be found for icaridine, isophorone and lauric acid. The derived PNEC values are <0.1 mg/L for DEET and isophorone and 0.3 mg/L for icaridine [27-30]. No conventional toxic effects on aquatic environments are expected for EBAAP [29].

One of the aspects to consider when suggesting suitable chemicals for experimental investigations is that testing material is available in sufficient amounts with high purity at reasonable prices. Preferably, reliable analytical methods should be in place, and quantitative standards should be commercially available. These preconditions are generally fulfilled for the candidate substances.

Comparative assessment of the available data revealed that there is no ideal substance with the potential to act as the one and only reference compound to test for possible anthropogenic infochemical effects on chemical communication in aquatic systems (Table 1). The three repellents DEET (134-62-3), icaridine (119515-38-7) and EBAAP (52304-36-6) complied best with the requirements. Isophorone, a natural attractant and at the same time an industrial chemical produced at more than 1,000 tonnes/year in Europe, may be an interesting fourth candidate. Isophorone is also used as solvent, just as was DEET. However, isophorone acts as attractant not as repellent [5], which would affect the experimental test design including the potential to act as positive control.

The data compiled in Table 1 for the four candidate substances show that they meet most of our criteria well,

with the exception of the degradability criterion where the heterogeneity of the data did not allow us to come to a consistent conclusion. The four selected candidate chemicals all have substantial infochemical potential due to the fact that smell is the active principle of their repellent or attractant effects. They have minor-to-moderate direct toxicities at low doses, and their physico-chemical properties indicate good handling under laboratory testing conditions. Furthermore, they are relevant to the environment since they occur in surface waters in considerable amounts. It is recognized though that modern repellents are readily degradable in sewage treatment plants. To make it very clear, it was not our intention to assess these chemicals *per se*. Instead, the data were only used for the ranking of substances to find promising candidates to detect infochemical effects under controlled laboratory settings.

Conclusions

We have identified potential test compounds with a high probability that their infochemical effects can be measured in suitable test systems. The selected four substances are odourants and act as attractants or repellents for target organisms, are moderately or non-toxic, are suitable for laboratory testing, and are relevant for the aquatic environment. The candidate chemicals are recommended for use in the subsequent part of the project, in which infochemical effects will be measured in established behavioural assays with invertebrates. Based on available data, we regard these substances as useful keys to open the first door to an understanding of infochemical effects from an ecotoxicological perspective.

The project consisting of the literature study presented here and the follow-up experimental work intends to contribute to the transfer of findings gained in chemical ecology to the practical assessment of infochemical effects in the field of ecotoxicology. This will be the first step to find out whether and to what extent anthropogenic substances might be responsible for infochemical effects on aquatic organisms. It also lays the groundwork for a strategy to identify potential infochemicals and to quantify their effects. This ambition can be regarded as comparable with the development of test and assessment strategies for endocrine disruptors when these effects were discovered two decades ago. The results obtained here with mostly biocides may also be useful for the analysis of the infochemical effect in other regulatory areas, such as pesticides, industrial chemicals and pharmaceuticals.

Competing interests

All authors declare that no competing interests exist, neither financial nor otherwise.

Authors' contributions

All authors (MN, UK and RB) drafted the manuscript together. MN did the data research. All authors read and approved the final manuscript.

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